

A UK framework for the assessment and integration of different scientific evidence streams in chemical risk assessmentt

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Title page

A UK Framework for the Assessment and Integration of Different Scientific Evidence Streams in Chemical Risk Assessment

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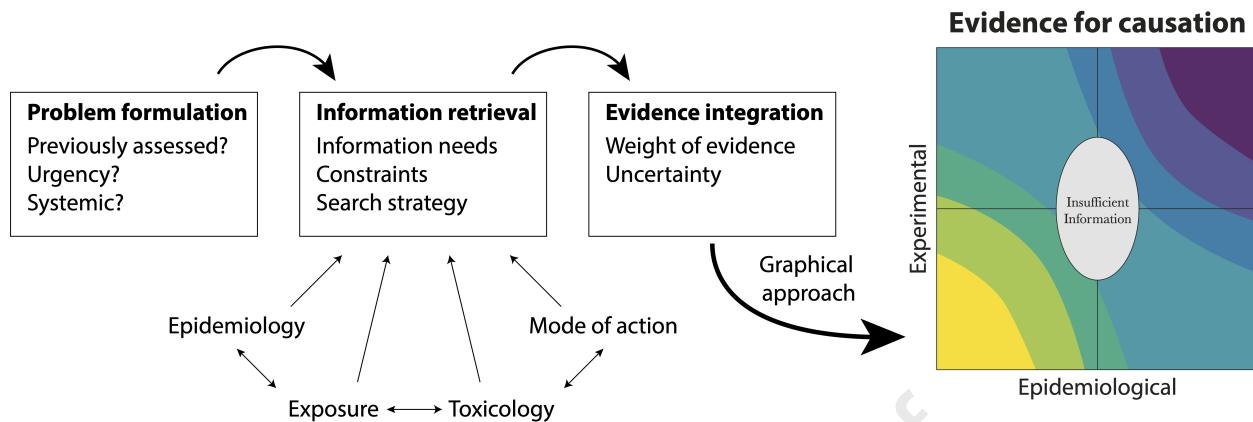
Abstract

Background: Few methods are available for transparently combining different evidence streams for chemical risk assessment to reach an integrated conclusion on the probability of causation. Hence, the UK Committees on Toxicity (COT) and on Carcinogenicity (COC) have reviewed current practice and developed guidance on how to achieve this in a transparent manner, using graphical visualisation.

Methods/Approach: All lines of evidence, including toxicological, epidemiological, new approach methodologies, and mode of action should be considered, taking account of their strengths/weaknesses in their relative weighting towards a conclusion on the probability of causation. A qualitative estimate of the probability of causation is plotted for each line of evidence and a combined estimate provided.

Discussion/Conclusions: Guidance is provided on integration of multiple lines of evidence for causation, based on current best practice. Qualitative estimates of probability for each line of evidence are plotted graphically. This ensures a deliberative, consensus conclusion on likelihood of causation is reached. It also ensures clear communication of the influence of the different lines of evidence on the overall conclusion on causality. Issues on which advice from the respective Committees is sought varies considerably, hence the guidance is designed to be sufficiently flexible to meet this need.

Synthesis and Integration of Epidemiological and Toxicological Evidence (SETE)



1 1. **Introduction/Background**

2 The assessment and integration of epidemiological, toxicological and other evidence
3 streams for risk assessment purposes is an integral part of the work conducted by
4 any scientific advisory committee (SAC). However, current approaches usually
5 consider epidemiological evidence separately from toxicological evidence, and then
6 combine the information at the end, most often in a non-systematic way. There are
7 several methods available for quantitative synthesis of epidemiological studies
8 (SEES, 2018; EFSA, 2020) but only a few methods exist for combining
9 epidemiological and toxicological studies to reach an integrated conclusion in a
10 transparent manner.

11 International bodies such as the International Programme of Chemical Safety
12 (IPCS), the European Food Safety Authority (EFSA) together with the Evidence-
13 Based Toxicology Collaboration (EBTC), the National Toxicology Program (NTP)
14 Office of Health Assessment and Translation (OHAT, Rooney et al. 2014), the
15 International Agency for Research on Cancer (IARC), the US Environmental
16 Protection Agency (EPA) and the Organisation for Economic Co-operation and
17 Development (OECD) have published guidance or frameworks which focus on or
18 include considerations on data integration in general, or for specific endpoints, e.g.
19 carcinogenicity. Several papers have also been published on the integration of
20 different evidence streams, focusing either on a general approach/framework (Adami
21 et al, 2011; Lavelle et al., 2012) or specific endpoints (Boyes et al., 2005) or
22 chemicals (Negri et al, 2017).

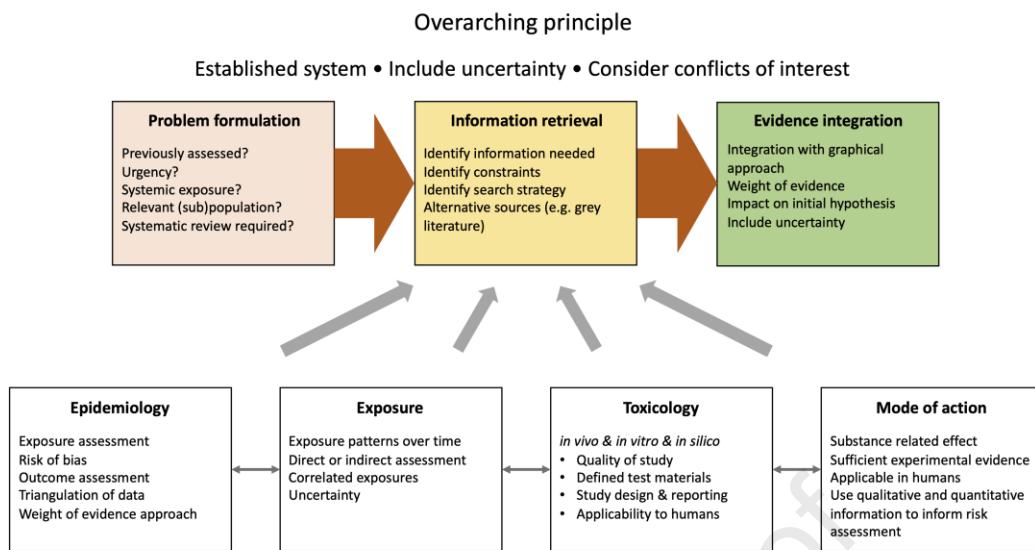
23 While all of the beforementioned frameworks and approaches have aspects or steps
24 in common, e.g. problem formulation, (systematic) literature reviews, and quality

25 assessment of studies, there are only a small number that provide practical and
26 applicable guidance on combining epidemiological and toxicological studies to reach
27 a conclusion on causality and none of these fully reflect the approach by the UK
28 SACs. Hence, in 2019, the Committee on Toxicity of Chemicals in Food, Consumer
29 Products and the Environment (COT) and the Committee on Carcinogenicity of
30 Chemicals in Food, Consumer Products and the Environment (COC) set up the
31 Synthesis and Integration of Epidemiological and Toxicological Evidence Subgroup
32 (SETE) to address this issue.

33 The aims of this paper are to briefly summarise the main considerations for the
34 assessment of different evidence streams, to provide pragmatic guidance and
35 transparent reflection on how the UK SACs review data, how different evidence
36 streams should be integrated in a transparent manner, giving appropriate weight to
37 all, and using graphical visualisation to ensure that the conclusions are explicit, and
38 clearly communicated. The paper thereby builds upon approaches for evidence
39 integration that have already been published (Adami et al., 2011; Lavelle et al., 2012;
40 Hart et al., 2010).

41 **2. Methods/Approach: Assessment and integration of different evidence
42 streams**

43 Detailed discussion of the quality assessment of the individual evidence streams and
44 a more in-depth discussion of the proposed evidence integration can be found in the
45 over-arching guidance of the SETE working group of the UK's independent COT
46 and COC (SETE, 2021). An overview of the approach developed is provided in
47 Figure 1 and key considerations are discussed in the following sections.



48

49 Figure 1: Overview of the key considerations for integrating different evidence
 50 streams, giving appropriate weight to each. Where possible, established systems
 51 should be used, and consideration should be given to uncertainties in the data.
 52 (Potential) conflicts of interest, especially where e.g. grey literature is used, should
 53 be clearly stated.

54 2.1 Problem Formulation and Information Retrieval

55 As a first (key) step, it is important to consider why a review or assessment is
 56 required, whether new information has become available, if a new potential risk has
 57 been identified, which population groups are to be addressed, and considerations
 58 whether individuals/groups could be at higher risk. This ensures that the right
 59 questions are asked, how urgently advice is needed, and helps make the most
 60 efficient use of resources and identifies the most appropriate approaches in a given
 61 situation. As information is retrieved and evaluated, the problem formulation may
 62 require refinement and additional aspects and considerations may be added. This
 63 should always be done in agreement with all relevant stakeholders.

64 It is important that the scope of the assessment is achievable and considers the
65 available resources. A systematic review is the optimal process to ensure all
66 available evidence has been identified and assessed, this is especially important the
67 greater the consequence of an issue or if the risk requires quantification. However,
68 an extensive systematic review is not necessary or possible in many situations and
69 e.g. recent systematic reviews available in the literature or by an authoritative body
70 can be utilised. Independent of the form of literature search, all studies relevant to
71 the endpoint in question, independently of the format, should be documented and
72 any changes to the initial search criteria should be recorded. All studies that provide
73 relevant data should be included at this point, bearing in mind that the process
74 begins with a specific question. However, the relevance and quality of studies will
75 need to be established by assessing, e.g. compliance with appropriate guidelines,
76 the relevance of the exposure, the nature of the adverse health outcome,
77 uncertainties and potential bias (SETE, 2021).

78 2.2 Considerations on different evidence streams

79 In assessing risks to human health from exposure to chemical substances, relevant
80 evidence/data comes from both experimental animals and human research, as well
81 as *in vitro* and *in silico* studies. Depending on the issue (e.g. risk from exposure to a
82 relatively new product), studies in experimental animals may provide the most
83 valuable, and perhaps even the only, information, whereas in other situations (e.g.
84 long-term and significant exposure to an environmental contaminant),
85 epidemiological studies may provide the most relevant information.

86 For both epidemiological and toxicological information, a weight of evidence (WoE)
87 approach should be applied, the specific details of the approach and

88 frameworks/guidance however may differ, depending on the information available. A
89 prescriptive, generic checklist or numerical scoring approach is not advised as such
90 an approach is likely to be limiting and inflexible. Epidemiological studies can provide
91 direct evidence of human health impacts of specific exposures and it is
92 recommended that, as far as possible, all relevant studies should be considered
93 (Lawlor et al., 2016). The combination of individual studies can provide strong
94 evidence, even if individually they may have different uncertainties and biases. *In*
95 *vivo*, *in vitro* or *in silico* toxicological studies have the ability to identify adverse health
96 effects of chemicals and provide mechanistic and experimental evidence for causal
97 associations, although human relevance is not always clear. The quality of each
98 study, using established criteria, for reliability, relevance and adequacy should be
99 assessed. Such studies can form the basis of estimating a concentration/dose likely
100 to be without appreciable effect in humans, if appropriate information is not available
101 from human studies. This approach is generally considered to be protective, but it
102 may (and indeed should) be modified if reliable scientific evidence is available
103 (Dourson et al., 1996).

104 A mode of action (MOA) and its associated key events provide a powerful bridge
105 between experimental studies (in animals, *in vitro* or *in silico*) and observations in
106 human populations. It underpins the weight of evidence considerations by providing
107 a mechanistic link between epidemiological observation and biological plausibility
108 (Boobis et al, 2006; 2008; Meek et al, 2014). Thus, an adverse effect in experimental
109 animals by a MOA that is considered relevant to humans would add appreciable
110 weight to the assessment of causality underlying an association with this outcome
111 observed in epidemiological studies, while a conclusion that a MOA is not relevant to

112 humans would argue against causality for the specific outcome in exposed subjects
113 (Boobis et al, 2006; 2008; Meek et al, 2014).

114 Increasingly, new approach methodologies (NAMs), comprising a range of *in vitro*
115 and *in silico* methods, are being used to assess the toxicological effects of
116 chemicals. However, the use of NAMs for regulatory decision making is still at an
117 early stage and hence current application is largely case-by-case. Guidance outlining
118 best practice for the development and implementation of NAMs for regulatory use in
119 human safety assessments are available (OECD, 2018; EURL ECVAM). However,
120 while methods which have undergone formal validation are robust, transferable and
121 widely trusted, the process of validation is time consuming and cannot keep pace
122 with the advances driving the development of NAMs. For NAMs to be widely
123 accepted in a future regulatory setting they need to be fit for purpose, with an
124 emphasis on methodological reliability/performance, biological/toxicological
125 relevance (e.g. linkage to key events) and interpretability for adverse effects *in vivo*.

126 Integrating data from NAMs with information from other sources, such as by
127 developing adverse outcome pathways (AOPs) or MOAs, can provide an additional
128 evidence stream in assessing qualitative and quantitative relationships between
129 adverse health effects and exposure in human populations. When considering
130 conclusions from new and yet to be validated, non-standard studies it is important to
131 assess the adequacy and relevance of the method as well as the results, especially
132 if a test system is far removed from humans (Kaltenhäuser et al., 2017).

133 Physiologically based pharmacokinetic (PBPK) modelling is particularly valuable for
134 the quantitative integration of data generated using *in vitro* and *in silico* methods and
135 may provide a means of bridging the exposure gap (Yoon et al., 2012). Studies with

136 unrealistic or unlikely exposure conditions for the general population may still provide
137 valuable insights into findings observed (or lack of) in epidemiological studies under
138 more relevant conditions. In assessing exposure, the emphasis is on assessing the
139 totality of the available information, which includes different sources and routes of
140 exposure, the assumptions and extrapolations made and the uncertainties that
141 remain in the resulting estimates. During evidence integration, the rationales, and
142 reasons for the choice of exposure information used for a given substance are
143 provided and the consequences and uncertainties of these choices for the overall
144 assessment are identified.

145 2.3 Evidence Integration

146 It is necessary to consider the overall picture when integrating evidence. No pre-
147 existing hierarchy for the different lines of evidence should be applied; however, it is
148 important to assess the confidence in the different lines of evidence and include this
149 in considerations of causality. Rarely is the process unequivocal, where all evidence
150 either supports or discounts a causal relationship. More often, information from
151 epidemiological and toxicological data is ambiguous and hence initially assessing the
152 strength of the lines of evidence separately will provide an indication of how reliable
153 that line of evidence is and in turn allows for an informed decision on how a specific
154 data set will influence the overall conclusion.

155 Building on previously published work as discussed above in Section 1, a number of
156 key points need to be considered when integrating epidemiological and toxicological
157 lines of evidence. These include whether a) the data indicate robust evidence of an
158 effect in animals and b) the same effect has been reported in epidemiological
159 studies. If the same effect has been reported in both animal and human studies,

160 consideration should be given as to how the effect levels compare. If possible,
161 internal concentrations should be compared, together with the relative sensitivities of
162 the molecular target and whether the effect concentration in the experimental studies
163 reflects a realistic exposure scenario in the general population. Furthermore,
164 consideration should be given to strain specific sensitivities to classes of
165 compounds. Information on AOPs or MOAs can further strengthen (or weaken) the
166 association between animal and human data and support for a biologically plausible
167 causal relationship. Considerations should be thereby given to whether there is
168 sufficient information to establish an MOE and whether the key events observed
169 experimentally occur in exposed humans. *In vitro* data can provide further support for
170 key events, if occurring at plausible concentrations, and are important to include in
171 the integration considerations, together with any other mechanistic data.

172 If a predominantly positive answer can be given to the main considerations, then the
173 WoE strongly supports causality. For example, *in vitro* data demonstrating that a key
174 event occurs at the same tissue concentrations as estimated in the exposed
175 population would add weight to a conclusion of causality, whereas the absence of
176 effects in occupationally or accidentally exposed populations at or above levels at
177 which effects are observed in experimental animals would reduce the weight of such
178 a conclusion. Consideration should also be given to whether a line of evidence is
179 considered sufficient by itself or provides a significant contribution to the overall
180 WoE.

181 Considerations of the lines of evidence, their strengths and weaknesses, and
182 specifically their influence on the conclusion should be clearly and transparently
183 stated. To assist discussion about the influence of different evidence streams on the

184 conclusion and causality, but also to allow for clear and easy communication a visual
185 representation of the conclusion of causality is recommended.

186 2.4 Constructing a Visual Representation for Causality

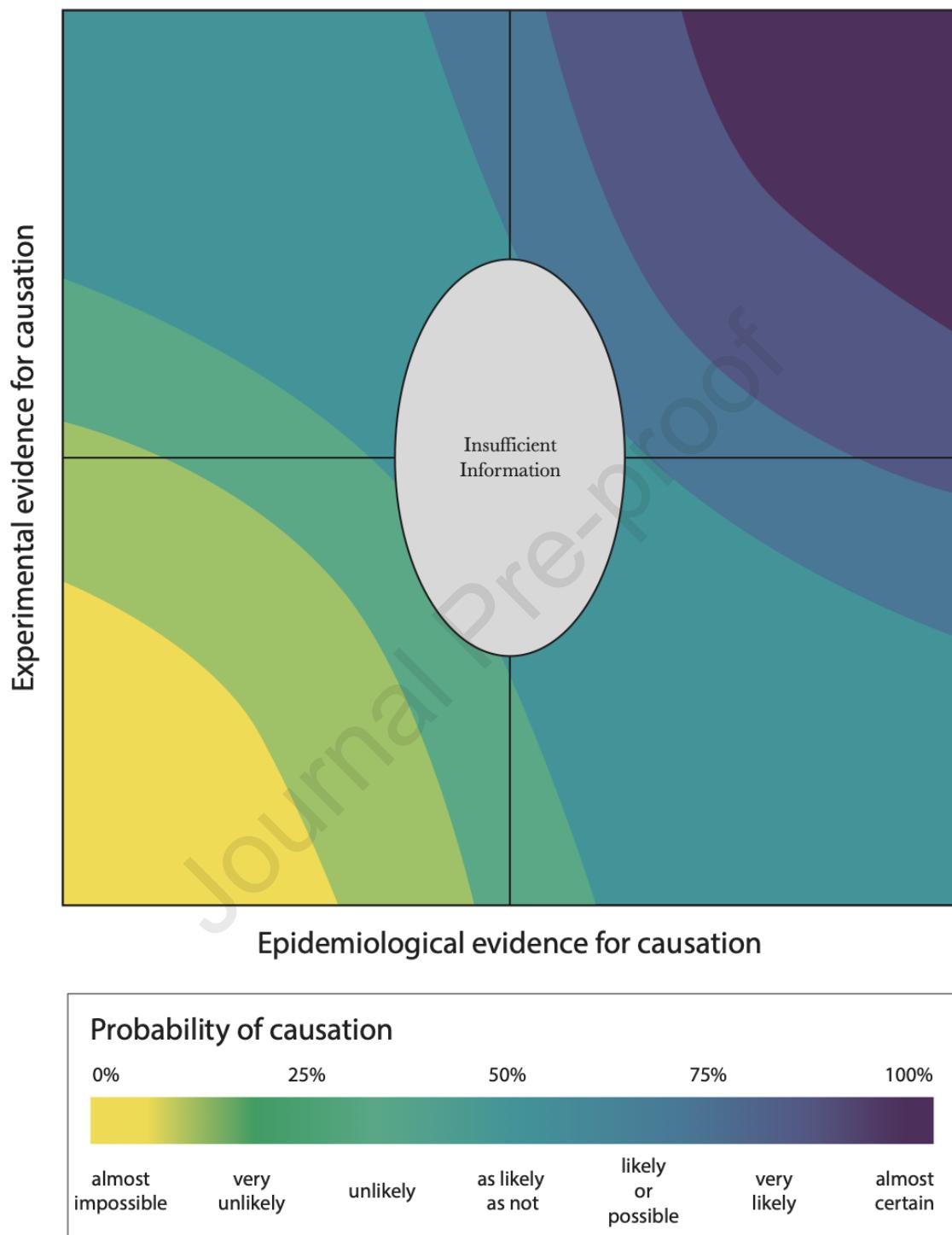
187 The visual representation, while a clear and easy way to communicate a conclusion
188 on causality, should always be accompanied by in depth discussion of the WOE and
189 underlying considerations by scientific experts.

190 Placement of the conclusion for a line of evidence on the probability of a causal
191 relationship on the graph is qualitative and is a deliberative process, based on the
192 considered professional judgment of the SAC. It requires assessment of all available
193 data and reflects what should already be current practice in chemical risk
194 assessment.

195 To support the construction of a visual representation it can be useful to establish a
196 line of evidence table summarising the strengths and weaknesses of the data as well
197 as the influence of the lines of evidence on the conclusion. This transparently
198 outlines the extent to which the data contribute to a conclusion on causality.

199 When producing the visual representation, it is important to start with a clear
200 hypothesis relating exposures to the substance of concern to adverse health effects
201 in humans. This forms the initial estimate of causal inference and should be placed
202 centrally in the grid. Depending on whether the toxicological, mechanistic or
203 epidemiological evidence previously assessed supports or discounts (or has no clear
204 influence on) a conclusion of causality, placement on the graph is moved
205 accordingly, either in a positive or negative direction. The movement itself is
206 influenced by the confidence in the initial estimate, using expert judgment, including

207 the impact of the strengths or weakness of the evidence, any relative weighing given
208 to epidemiological and toxicological studies and the uncertainties associated with the
209 data. Where possible estimates of uncertainty should be included, e.g. likely, upper
210 and lower bound of impact. The final positioning on the graph should reflect the
211 Committee's agreed conclusions on the weight of evidence on the likelihood of
212 causation, i.e whether a causal relationship is likely/unlikely, possible but lacks
213 strong experimental or epidemiological support or the information is insufficient to
214 reach a conclusion. An example of such a visual representation is provided in Figure
215 2.



216

217 Figure 2: Example for the visual representation of the likelihood of a causal
 218 relationship, considering both epidemiological and experimental data. Causality and

219 placement on the graph are qualitative and based on professional judgment of the
220 whole database.

221 The colour scheme and presentation of probability follows the UK PHIA framework,
222 or probability yardstick. In contrast to other approaches, the axes should not be
223 considered numerical, and it is not intended that there is a quantitative relationship
224 between increments along an axis. Instead, positioning on the graph is the result of a
225 deliberative process and reflects the increasing or decreasing WoE based on expert
226 judgment on the likelihood (or not) of causation from exposure to a chemical leading
227 to an adverse outcome in exposed populations. Rather than a probabilistic or
228 numerical approach, the above visualisation is intended as a transparent means of
229 communicating the agreed conclusions. The final conclusion of the assessment
230 should be stated, with an estimate of the overall uncertainty and, where appropriate,
231 guidance on how data gaps could be filled.

232 While assessments of different evidence streams are often lengthy undertakings, as
233 more information is included in the process and/or becomes available, the placement
234 of the experimental and/or epidemiological evidence on the graph can be easily
235 adjusted.

236 **2.4.1 Example of evidence integration**

237 Cadmium, a contaminant with a well-established adverse effect, nephrotoxicity, was
238 chosen to illustrate the principles and considerations of the SETE guidance on
239 evidence integration. No full assessment of cadmium was undertaken but rather the
240 lines of evidence were drawn from previously published assessments (EFSA, 2009;
241 2011) and analysed for how these impacted on the WoE for a causal relationship
242 between cadmium exposure and nephrotoxicity. It should be stressed that the

243 following assessment is for illustrative purposes only; a full assessment would
 244 require a much more deliberative process, including a comprehensive problem
 245 formulation and WoE assessment.

246 Cadmium, in brief, primarily affects the kidney, especially the proximal tubular cells,
 247 where it accumulates and may cause renal dysfunction. Cadmium can also cause
 248 bone demineralisation (directly through bone damage or indirectly through renal
 249 dysfunction). After prolonged and/or high exposure tubular damage may progress to
 250 decreased glomerular filtration rate and eventually renal failure. It should be noted
 251 that both EFSA (2009; 2011) and JECFA (FAO/WHO, 2011) identified renal toxicity
 252 as the critical effect for establishing a health-based guidance value for cadmium. The
 253 example presented here focused on nephrotoxicity.

254 The target organ (kidney) and the toxicokinetics after oral exposure are similar
 255 among species, however the estimated absorption of cadmium in rodents is lower
 256 compared to humans, especially after prolonged exposure. In addition, some species
 257 differences in metallothionein synthesis, to which cadmium binds, cadmium kinetics
 258 and toxicity have been well established.

Lines of evidence and their main strengths (S) and weaknesses (W)	Influence on Conclusion
Animal data S – The target organ (kidney) and the toxicokinetics after oral exposure are similar among species (including humans)	While there are species specific differences in metallothionein expression, cadmium kinetics and toxicity, these differences are well established and the animal data (target organs/endpoints) are in support of human findings

<p>S – Cadmium is a clear nephrotoxin in experimental studies</p> <p>W – Estimated absorption of cadmium in rodents is lower compared to humans, especially after prolonged exposure</p>	
<p>Human data</p> <p>S – Consistent evidence that cadmium targets kidney after chronic exposure</p> <p>S – While renal toxicity is not as evident at low exposures, there is a clear indication of a positive dose-response relationship</p> <p>W – Results of cross-sectional studies affected by some degree of imprecision, which could cause an underestimation of true cadmium toxicity</p> <p>W – No firm conclusion on reversibility of renal damage, some data indicate possibility, others note glomerular dysfunction to progress even after contaminated soil replacement</p>	<p>Strong evidence that cadmium is a nephrotoxin from epidemiological studies and environmental exposure</p>
<p>Mechanistic data</p> <p>S – Link between the MOA, key events and human data</p>	

Conclusions on causality	Epidemiological and experimental animal data and information on MOE provide strong evidence for a causal relationship between exposure to cadmium and renal toxicity.
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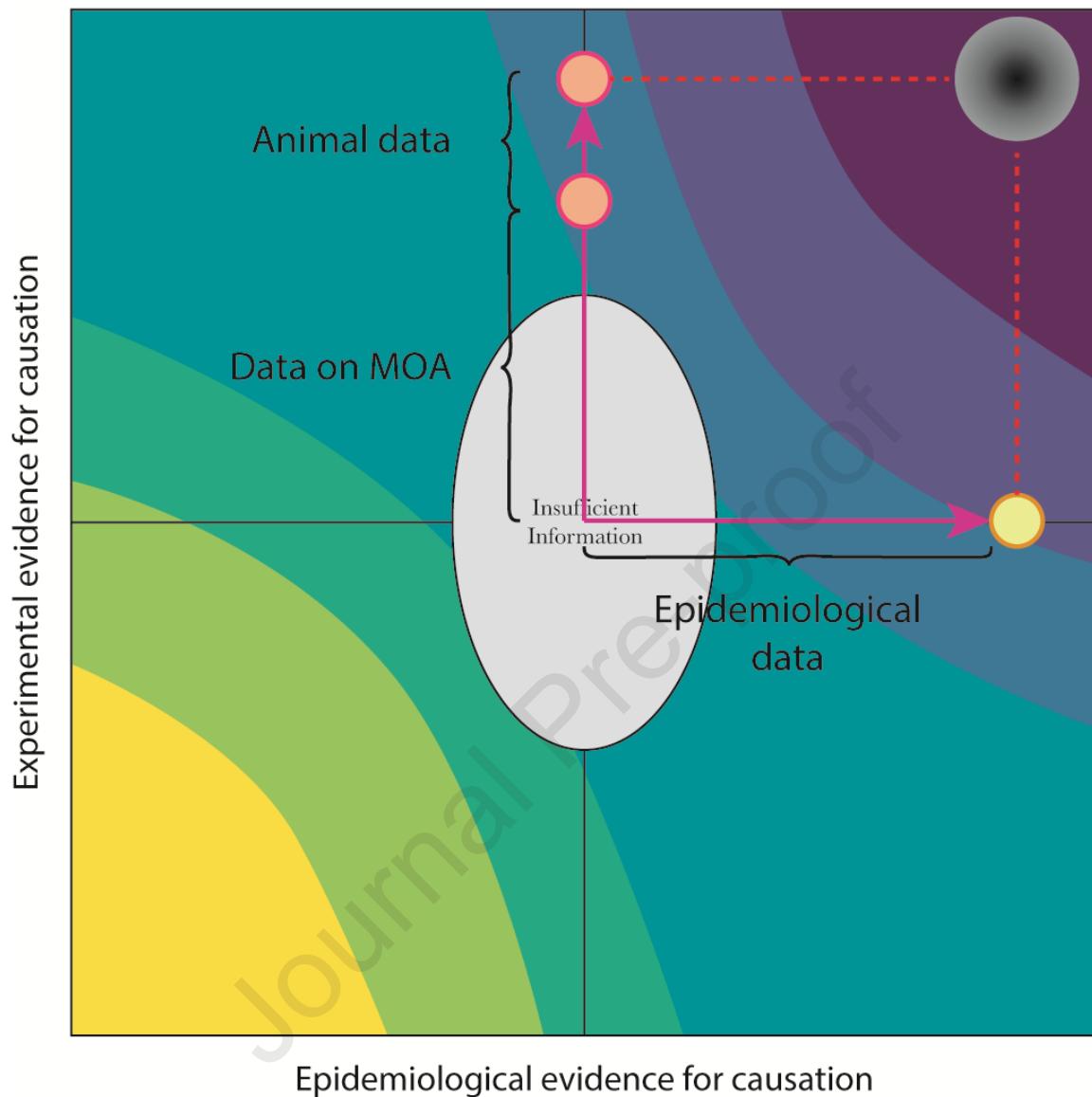
259

260 Table 1: Summary of the strengths and weaknesses of the data on cadmium and the
 261 influence of the lines of evidence on the overall conclusion. Please note the lines of
 262 evidence and conclusions on the strengths and weaknesses have been drawn from
 263 previous evaluations and have not been systematically assessed here.

264 The available epidemiological studies provide consistent evidence that cadmium
 265 causes renal damage in some human populations. While the effect at low exposures
 266 is not as apparent, a positive dose-response relationship can be clearly identified,
 267 with increasing effect at increasing doses. Renal toxicity has been reported in
 268 epidemiological studies considering not only occupational exposures but also after
 269 environmental exposure or exposure through drinking water. The renal effect in
 270 humans is further supported by animal data, identifying cadmium as a classic
 271 nephrotoxin. While there are some species differences, specifically in
 272 metallothionein, cadmium kinetics and toxicity, these differences are well established
 273 and the animal data, i.e. target organs/endpoints, are in support of human findings.
 274 Both, epidemiological and experimental animal data provide strong evidence for a
 275 causal relationship between cadmium exposure and nephrotoxicity in humans. This
 276 is further supported by mechanistic data, providing a link between the MOA and
 277 human data.

278 For the recommended visualisation, the conclusion of a strong association of
279 cadmium exposure and nephrotoxicity was applied. Starting in the middle of the
280 graph and given the strong epidemiological evidence for such an association the
281 marker was set to the far right. Both animal data and mechanistic data, here the
282 MOA, also provide strong evidence for a causal association, hence the second
283 marker was set near the top of the axis (again starting from the middle). The final
284 conclusion on causality is visualised where the two lines intersect, and the final
285 marker is placed. In this example, a causal association between cadmium exposure
286 and nephrotoxicity, based on the consideration and integration of all available
287 evidence, is almost certain.

Cadmium



288

289 Figure 3: Visualisation of the likelihood for a causal relationship between cadmium
 290 exposure and nephrotoxicity. The yellow circle is representative of all
 291 epidemiological evidence assessed; the upper orange circle of all toxicological

292 evidence assessed. The lower orange circle indicates the impact of evidence on
 293 MOA on the conclusions. As all lines of evidence strongly suggest an effect, they
 294 have been moved to a place at the top (experimental) and far right (epidemiological).
 295 The grey circle represents the conclusion on causality from integration of all of the
 296 evidence and has been set where the individual lines of evidence intersect. Causality
 297 and placement on the graph are qualitative and based on professional judgment of
 298 the whole database.

299 **3. Discussion and Conclusions**

300 The aims of this paper are to build upon published approaches for evidence
 301 integration (Adami et al., 2011; Lavelle et al., 2012; Hart et al., 2010) and provide
 302 pragmatic guidance and transparent reflection on how the UK SACs review data and
 303 how different evidence streams should be integrated in a transparent manner, using
 304 graphical visualisation, giving appropriate weight to all.

305 Some work on how to integrate different evidence streams has been conducted at an
 306 international level. While existing approaches have certain aspects or steps in
 307 common, in general, they do not provide applicable and transparent guidance on
 308 how the actual evidence integration is/or should be undertaken. While the work here
 309 includes considerations on the same steps as in other approaches, i.e. problem
 310 formulation, literature retrieval and the assessment of the different evidence streams,
 311 by using established systems, the main focus is on the integration of different data
 312 sets and their visual representation. When integrating evidence, all lines of evidence
 313 should be considered, with no pre-existing hierarchy.

314 Good risk assessment practice involves a transparent description of consideration of
 315 the relevance of the endpoint(s) and adverse effects in/to human exposure, i.e. do

316 the data indicate a causal relationship, based on robust evidence of an effect in
317 animals and has the same effect been reported in epidemiological studies, as well as
318 whether the effect concentration in animals is of biological relevance in the general
319 population. Consideration should also be given to whether mechanistic data such as
320 information on AOPs or MOAs, are available as they can further strengthen (or
321 weaken) the support for a biologically plausible relationship. Information should also
322 be provided on potential biases/uncertainties in the data.

323 While this paper only briefly summarises the key aspects to be considered, reflecting
324 what should be current practice, further details are provided in the COT's and COC's
325 2021 SETE report (specifically Annex 1).

326 The novel aspect of this paper is the inclusion of a visual representation of the
327 conclusion on causality. This requires an explicit conclusion on the qualitative
328 contribution of the different lines of evidence on the probability of a causal
329 relationship between (usually) a specific adverse effect and exposure of the
330 population. The requirement to place a representation of this on the graph, means
331 that experts will need to agree a conclusion on the totality of the available data on a
332 line of evidence. The need to plot the different lines of evidence on the same graph
333 requires appropriate weighing of their relative contribution to the probability of
334 causation. Hence, visual representation provides a means for improving the
335 transparency and clarity of discussions on causation.

336 In addition, visual representation not only facilitates simple and clear communication
337 of the SAC's conclusion on causality, but also the influence that the different
338 evidence streams have on the final conclusion. This can help identify evidence gaps
339 and research needs. While the scale used for visualisation of probability follows a UK

340 government established system for communication of probability, in the scheme
341 proposed here, conclusions on causality are qualitative, based on expert judgment,
342 not quantitative. The further along the axes the circle is set, the more weight the data
343 has been given in supporting a causal relationship. Where the different evidence
344 streams intersect, the conclusion on causality is easily depicted and is a simple
345 means of clearly indicating a consensus view. Again, the authors would like to stress
346 that the movement on the graph is based on expert judgement and that the
347 visualisation should always be accompanied by a detailed assessment of the
348 underlying data.

349 Integration of information derived from epidemiological and toxicological studies
350 requires an appreciation of the scientific processes around different disciplines to
351 allow for an appropriate and balanced, evidence-based conclusion regarding
352 causality. Ongoing communication among experts in the different disciplines is
353 therefore essential to ensure a shared understanding of the question(s) to be
354 addressed and the planned outputs of the risk assessment or other advice/evidence.
355 This overview provides an approach and a practical example of how such integration
356 can be applied successfully.

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Highlights

- Approach to improve the consistency, transparency and communication of current practice of scientific advisory committees
- Emphasis is on weight of evidence and integration of different evidence streams, where all evidence is considered
- Visualisation tool is proposed, to help communicate the overall conclusion and contribution of different lines of evidence
- The conclusion on causality and its graphical representation is qualitative, and based on expert judgement
- Collaboration and ongoing dialogue between the different disciplines in risk assessment are strongly advised